

# Uncertainty of Measurement from sampling, preparation and chemical analysis

Michael H Ramsey FRSC

School of Life Sciences,  
University of Sussex, Brighton, UK  
m.h.ramsey@sussex.ac.uk

Uncertainty in Contaminated  
Land Risk Assessment. RSC-Tox  
SOBRA, Burlington House  
London, 17 December 2009



## Overview

- Uncertainty (U) in general (e.g. in Risk Assessment)
- Uncertainty of Measurements (UoM)
  - Measurement process – includes sampling & prep
- Methods for estimating uncertainty of measurements
  - e.g. Guidance from Eurachem/Eurolab/Citac/Nordtest/RSC-AMC
- Benefits of knowing uncertainty
- Conclusions



## Uncertainty in general

- Several types of uncertainty ranging from determinism and total ignorance (adapted from Walker *et al.*, 2003)



- Separate concepts of **variability** (spread of true values), from **uncertainty** (U) (Kelly & Cambell, 2000)
- U - includes all contributions to the estimated (rather than the true) distribution of values, that come from imperfect knowledge.
- **Uncertainty** can be reduced, e.g.:-
  - U of measurement - by taking larger sample mass
  - U-of-mean value - by repeated measurements
 – but **variability** cannot be reduced.
- Statistical uncertainty – is the most quantitative form, one example is...
- Uncertainty of measurement

Walker, W.E., Harremoes, P., Rotmans, J., van der Sluis, J.P., van Asselt, M.B.A., Janssen, P. and von Krauss, M.P.R. 2003 Defining uncertainty. A conceptual basis for uncertainty management in model-based decision support. *Integrated Assessment* 6, 5-17

Kelly, E.J., Campbell, K. (2000) Separating variability and uncertainty in environmental risk assessment - Making choices. *Human and Ecological Risk Assessment* 6 (1): 1-13



## Uncertainty of measurement (inc. sampling)

- U of measurement is:-
  - **Informally**:- the interval around the result of the measurement that contains the **true value** with high probability
  - **Formally**:-
    - An estimate attached to a test result which characterises the range of values within which the true value is asserted to lie (ISO, 1993<sup>1</sup>)
    - Non-negative parameter characterizing the dispersion of the quantity values being attributed to a measurand, based on the information used (VIM, 2008<sup>2</sup>)
  - Includes random and systematic effects. U ≠ precision
  - Ideally U value attached to **each measurement**  $x \pm U$ 
    - Gives user info on quality (not left in the lab!)
- U arises from **all** steps in measurement (e.g. sampling & prep)
- Key parameter of measurement (and sampling) quality
  - Encompasses effects of other parameters of analytical **methods**,
    - e.g. bias, precision, detection limit, etc.
- Doesn't assume measurements (or sampling) are 'correct'

<sup>1</sup>ISO 3534-1: 1993 Statistics – Vocabulary and Symbols, International Organization for Standardization, Geneva

<sup>2</sup>International Vocabulary of Metrology – Basic and General Concepts and Associated Terms. VIM, 3rd edition, JCGM 200:2008



## Uncertainty of Measurement (UoM)

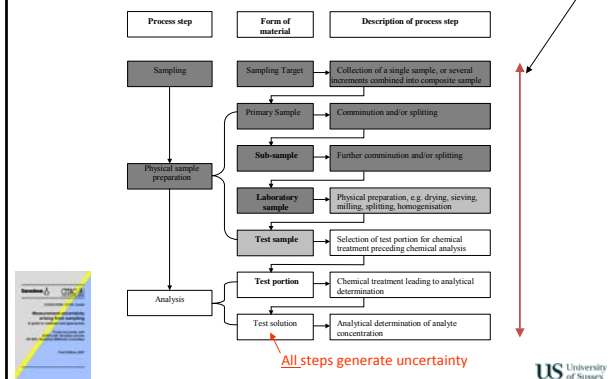
- The interval around the result of a measurement that contains the true value with high probability  
e.g.  $U = 20\%$  on measurement of 390 mg/kg Pb
  - Range of  $U = 390 + 20\% = 468$  mg/kg
  - to  $390 - 20\% = 312$  mg/kg
- We can never know the true value (could be 460 mg/kg)
- Effect on comparison with threshold value of 450 mg/kg  
i.e. Measured value 390 mg/kg is **below** the threshold value  
True value 460 mg/kg is **above** the threshold value  
Measured value gives 'false negative' classification
- All we need to know is how far from the truth we might be
  - e.g. with 19/20 chance if being right = 95% confidence
- Different from U on mean value (used in CIEH/CLAIRE Guidance)
- How can we estimate uncertainty of measurements
  - including that from sampling?
  - Eurachem Guide\* has methods for estimating measurement uncertainty
    - including that arising from sampling
    - Has applications to soil, groundwater, food, gas



\*Ramsey M.H., and Ellison S. L. R., eds. (2007) Eurachem/EUROLAB/ CITAC/Nordtest/ AMC Guide: Measurement uncertainty arising from sampling: a guide to methods and approaches Eurachem ISBN 978 0 948926 26 6. ([http://www.eurachem.org/guides/US\\_2007.pdf](http://www.eurachem.org/guides/US_2007.pdf))



## Measurement Process – Starts with sampling



## Four empirical methods for UoM including that from sampling

| Method # | Method description | Samplers (People) | Protocols | Component estimated                 |               |                 |                  |
|----------|--------------------|-------------------|-----------|-------------------------------------|---------------|-----------------|------------------|
|          |                    |                   |           | Sampling Precision                  | Sampling Bias | Anal. Precision | Anal. Bias       |
| 1        | Duplicates         | single            | single    | Yes                                 | No            | Yes             | No <sup>1</sup>  |
| 2        | Multiple protocols | single            | multiple  | between protocols                   |               | Yes             | No <sup>1</sup>  |
| 3        | CTS                | multiple          | single    | between samplers                    |               | Yes             | Yes <sup>2</sup> |
| 4        | SPT                | multiple          | multiple  | between protocols +between samplers |               | Yes             | Yes <sup>2</sup> |

CTS = Collaborative Trial in Sampling, and SPT = Sampling Proficiency Test.

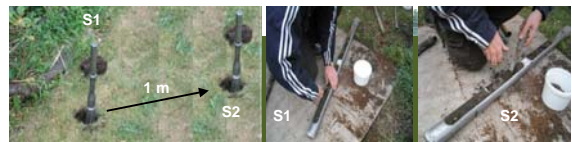
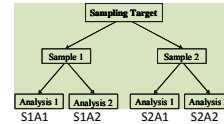
Simplest Empirical method is 'Duplicate Method' (#1)

- 1 Estimate analytical bias using CRM (Certified Reference Materials).
- 2 Analytical bias partially or completely included where multiple labs involved



## The Duplicate Method

- Define your sampling target and sampling protocol (Exploratory or Main)
- Use a balanced design



- Take a sample at the nominal sampling target.
- At 10% of locations (n≥8) - take a second sample displaced from the original (in space or time) to reflect the ambiguity in the sampling protocol
- Carry out duplicate analyses on both the sample duplicates
- Estimate uncertainty components using Robust ANOVA



## Case Study #1: Duplicate Method for estimation of UoM in hazard assessment

The scenario:

- 9 hectare on former landfill, in West London
- Potential housing development
- key contaminant → Pb
- Need to delineate area areas >SGV
- Prior to site-specific risk assessment
- 100 sampling locations – in regular sampling grid, 10 x 10, spacing 30 m →
- top soil samples (0 – 0.15 m) - collected using a sampling auger
- survey conducted with measuring tape and compass
  - U in relocation ~ 3m ( - similar to that using typical GPS)



Eurachem UFS Guide: Example A2



## Study design – Duplicate Method

- Duplicate samples taken at 10 sampling locations (i.e. 10%) randomly selected.
- 3 m from the original in a random direction
- Aims to reflect :-
  - the ambiguity in the sampling protocol
  - the uncertainty in locating the sampling target (e.g. survey error)
  - the effect of small-scale heterogeneity on the measured concentration within the sampling target



## Sample prep and analysis in the lab

- 6 soil CRMs were selected for analysis to assess the analytical bias over a range of concentrations
- Measurements subject to full AQC – but not assumed to be 'true' values of concentration
- Corrected for reagent blank concentrations where statistically different to zero
- Raw measurements for estimation of uncertainty were:
  - unrounded – e.g. 0.0124 mg/kg not < 0.1 or < detection limit
  - unrounded – e.g. 2.64862 mg/kg not 3 mg/kg



## Results – map of Pb contamination

| Row | A   | B    | C   | D   | E   | F    | G   | H   | I   | J   |
|-----|-----|------|-----|-----|-----|------|-----|-----|-----|-----|
| 1   | 474 | 287  | 250 | 338 | 212 | 458  | 313 | 125 | 77  | 168 |
| 2   | 378 | 3500 | 260 | 152 | 197 | 711  | 165 | 69  | 206 | 126 |
| 3   | 327 | 197  | 240 | 159 | 327 | 264  | 105 | 137 | 131 | 102 |
| 4   | 287 | 207  | 197 | 87  | 254 | 1840 | 78  | 102 | 71  | 107 |
| 5   | 395 | 165  | 188 | 344 | 314 | 302  | 284 | 89  | 87  | 83  |
| 6   | 453 | 371  | 155 | 462 | 258 | 245  | 237 | 173 | 152 | 83  |
| 7   | 72  | 470  | 194 | 83  | 162 | 441  | 199 | 326 | 290 | 164 |
| 8   | 71  | 101  | 108 | 321 | 218 | 327  | 340 | 132 | 258 | 246 |
| 9   | 72  | 188  | 104 | 463 | 482 | 228  | 135 | 285 | 181 | 146 |
| 10  | 89  | 366  | 495 | 779 | 60  | 206  | 56  | 135 | 137 | 149 |

Argyaki (1997)

Only 16/100 locations over UK SGV = 450 mg Pb/kg

— mainly 'uncontaminated' = below action limit (84%)

- US<sub>95</sub> (235 mg Pb/kg) < 450 mg Pb/kg – no action

- The max value at location B2 (3590 mg Pb/kg) was shown to be a population outlier



## Results from Duplicate Method

- Low level agreement between sample duplicates (e.g. D9) high level of sampling uncertainty
- Agreement between analytical duplicates much better < 10 % difference
- Robust ANOVA selected to allow for the outlying values evident in this data.

| SAMPLE I.D. | S1A1 (mg kg <sup>-1</sup> ) | S1A2 (mg kg <sup>-1</sup> ) | S2A1 (mg kg <sup>-1</sup> ) | S2A2 (mg kg <sup>-1</sup> ) |
|-------------|-----------------------------|-----------------------------|-----------------------------|-----------------------------|
| A4          | 787                         | 769                         | 811                         | 780                         |
| B7          | 338                         | 327                         | 651                         | 563                         |
| C1          | 289                         | 297                         | 211                         | 204                         |
| E8          | 229                         | 215                         | 208                         | 218                         |
| F7          | 346                         | 374                         | 525                         | 520                         |
| H5          | 56                          | 61                          | 77                          | 73                          |
| I9          | 189                         | 189                         | 176                         | 168                         |
| J5          | 61                          | 61                          | 91                          | 119                         |

## Components of UoM Random & Systematic

|              | $s = u$<br>mg kg <sup>-1</sup> | $U = ks$<br>mg kg <sup>-1</sup> | $U_{meas} / \bar{x}$<br>= $\frac{200 \cdot s_{meas}}{\bar{x}}$ | % of $U_{meas}$ |
|--------------|--------------------------------|---------------------------------|--|-----------------|
| Analysis     | 11.1                           | 22.2                            | 7.5%   | 0.8%            |
| Sampling     | 123.8                          | 247.6                           | 83.3%  | 99.2%           |
| Measurement* | 124.3                          | 248.6                           | 83.6%  | 100%            |

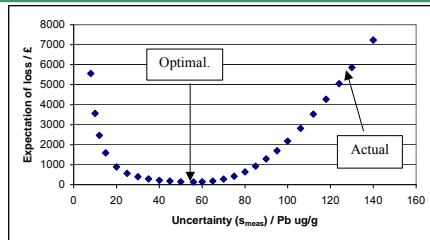
$$*U_{meas} = s_{meas} = \sqrt{(s_{samp}^2 + s_{anal}^2)} = 124.3 \text{ mg kg}^{-1}$$

- averaged over the 10 targets
- to represent the 100 targets
- Expressed in relative terms (%) for widest applicability
- Systematic error – from Analytical Bias estimated as 3.41 % ± 1.34 %
- By analysis of 6 CRMs – bias modelled as function of concentration
- $U_{systematic} = 2 \cdot \sqrt{(-3.41^2 + 1.34^2)} = 7.33 \%$
- Adding in this uncertainty
- $U_{total} = \sqrt{(U_{random}^2 + U_{systematic}^2)} = \sqrt{(83.63^2 + 7.33^2)} = 83.95$
- Effectively identical to estimate based on just random components
- Does exclude bias from primary sampling – other methods required

## If site mean is the measurand?

- If the measurand (or true value) had been defined as the mean concentration across the whole site
- the uncertainty estimate should include the contribution of the standard error on the mean
- In this case study:
  - $s_{total} = 403 \text{ mg kg}^{-1}, \bar{x} = 291.9, n = 100$
  - $se = \frac{s_{total}}{\sqrt{n}} = \frac{403}{\sqrt{100}} = 40.3 \text{ mg kg}^{-1}$
  - the relative expanded uncertainty on the mean is:
    - $= 200 \frac{se}{\bar{x}} = 200 \times \frac{40.3}{291.9} = 27.6 \%$  of the mean value

## Optimal Uncertainty vs Actual for West London Case Study

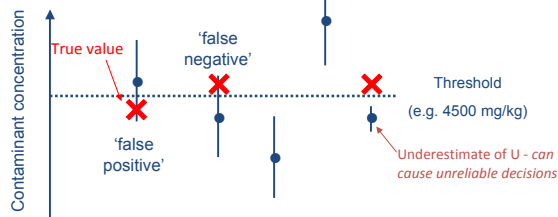


- False Positive at  $c = 700 \mu\text{g g}^{-1}$  unnecessary remediation

- Loss predicted at actual uncertainty £5680 per location
- £130 at optimised uncertainty - save 96%
- 3 fold reduction in  $U_{samp}$  indicated (e.g. by taking 9-fold composite samples)
- Judge Fitness-for-purpose of measurements – (not use labs criteria e.g. 10%)

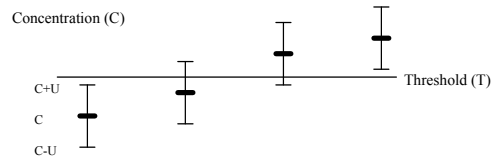
Ramsey et al. (2002) J. Env. Monitoring, 4, 5, 809 - 814

## Benefits of Knowing UoM → make more reliable decisions



## Probabilistic classification

- Should always report the uncertainty values with the analyte concentration value as it can alter the classification of the sampling location.
- Allows a probabilistic classification:



Definitely under the action level | Possibly over the action level | Probably over the action level | Definitely over the action level | Probabilistic Classification

## Probabilistic classification- Mapping

| Row | A   | B    | C   | D   | E   | F    | G   | H   | I   | J   |
|-----|-----|------|-----|-----|-----|------|-----|-----|-----|-----|
| 1   | 474 | 287  | 250 | 338 | 212 | 458  | 713 | 125 | 77  | 168 |
| 2   | 378 | 3590 | 280 | 152 | 197 | 711  | 165 | 69  | 206 | 126 |
| 3   | 327 | 197  | 240 | 159 | 327 | 264  | 105 | 137 | 131 | 102 |
| 4   | 787 | 207  | 197 | 87  | 254 | 1840 | 78  | 102 | 71  | 107 |
| 5   | 395 | 165  | 188 | 344 | 314 | 302  | 284 | 89  | 87  | 83  |
| 6   | 453 | 371  | 155 | 462 | 258 | 245  | 237 | 173 | 152 | 83  |
| 7   | 72  | 470  | 194 | 83  | 162 | 441  | 199 | 326 | 290 | 164 |
| 8   | 71  | 101  | 108 | 521 | 218 | 327  | 540 | 132 | 258 | 246 |
| 9   | 72  | 188  | 104 | 463 | 482 | 228  | 135 | 285 | 181 | 146 |
| 10  | 89  | 366  | 495 | 779 | 60  | 206  | 56  | 135 | 137 | 149 |

41% of locations at least 'possibly over the action limit'

## Case Study 2. Metal working site

- Former industrial landfill site in the UK.
- approx. 1.68Ha area
- Used for waste disposal for approximately 40 years.
- Exact nature of the wastes deposited at the site are unknown
- Suspected to contain elevated concentrations of heavy metal and organic (potentially diesel, hydraulic, rolling and heavy fuel oils and lubricating greases) contaminants.
- Experiment designed for comparison of on-site versus lab-base measurement uncertainty



Boon R.A., Ramsey M.H., McIntyre S., and Yeo M. (2008) The use of measurement uncertainty to assess the reliability of on-site field test kits for the investigation of contaminated land. Proceedings of ConSoil 2008 (10<sup>th</sup> International UFZ-Deltras/FNO Conference on Soil-Water Systems), Milan, Italy, 3-6 June 2008 (ISBN: 978-3-00-024558-5), Theme C, 64-73

## Case Study 2 - Methods

### - field sampling and preparation



- 11 trial pits in stratified random sampling design
  - Excavated using mechanical digger

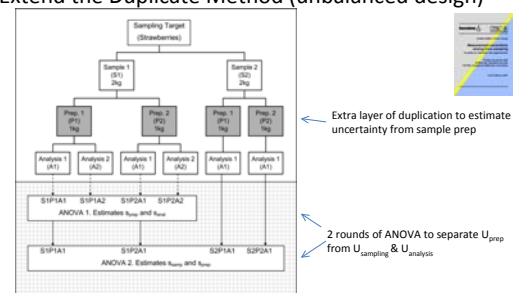


- Primary sample (4-fold composite) collected from pile of soil at side of pit for each 0.5 m interval.
- Total of 62 samples were collected.
- Duplicate Method (+CRMs) used to estimate U-
  - Duplicate samples at 8 sampling locations
  - by reinterpreting the sampling protocol.
- Primary sample prepared in the field
- Measurements both in field (and in lab), for TPH (Total Petrol Hydrocarbons) and others (PAH, metals)
- Extra duplicates to estimate UoM from sample preparation
  - Design from Eurachem Guide



## UoM from Sample Preparation

- Extend the Duplicate Method (unbalanced design)



Eurachem UFS Guide (2009) Appendix D, and Lyr, J.A., Ramsey, M.H., Fusse, R.J. and Wood, R. (2003) Measurement uncertainty from physical sample preparation: estimation including systematic error. Analyst, 2003, 128 (11), 1391 - 1398

## Sample Prep Contribution to UoM on Field measurement of TPH

|          | s (mg/kg) | U(%) | % of UoM |
|----------|-----------|------|----------|
| Sampling | 173.0     | 32   | 49       |
| Prep     | 125.4     | 32   | 26       |
| Analysis | 124.0     | 23   | 25       |

Combined 'U from Sampling' = 75%



- Using modified Duplicate Method - with extra sample prep duplicates
- Robust ANOVA to separate & quantify variance
- Show that U from 'Sampling' (as usually defined) contributes 75% of UoM
- ~35% of this arises from Sample Preparation = substantial - in this case
- Lab measurements of TPH also high U ( $U_{anal} = 32\%$ )
  - even though accredited method (lab's estimate of  $U_{anal}$  much lower ~ 15%)
  - Lab sample prep also hidden source of uncertainty - not reported

## Conclusions (1)

- Uncertainty arises at every step in risk assessment,
  - but Uncertainty of Measurement (UoM) is often overlooked
- UoM can affect both the construction of RA models
- and its interpretation = double effect
- UoM is not negligible just because measurements are made in a 'accredited' lab
  - as analytical uncertainty can be unexpectedly high in reality
  - Most uncertainty arise from the sampling and sample preparation (UFS not quoted by the lab)
  - dominates the budget of measurement uncertainty
- Methods are now available to estimate UoM (inc UFS) (Eurachem Guide)
  - applicable to many media such as soil, water, air, wastes, food, feed etc.
  - New Support Tool for Reliable Environmental Measurement (STREM) includes methodologies

## Conclusions (2)

- New methods now revealing the size of the UoM for the first time
  - examples in Cont Land often 50 - >150%
  - extra information that adds to *U of mean value* used in ClEH/CLAIRE Guidance
- Knowing UoM enables:-
  - Propagation of UoM into Human Health Risk Assessment
    - random component can be overwhelmed by other sources (e.g. dose-response),
    - but systematic component remains
    - hence more reliable RA and environmental management decisions
    - example for Blood Lead published\*
  - Ability to compare different sources of UoM, see which is limiting
    - decide whether to reduce that sources
      - e.g. take samples with larger mass or more increments
    - Don't rely on labs estimate of  $U_{optimal}$
  - Enable the concept of 'optimal' levels of UoM = Fitness-for-purpose of measurements
    - decided by risk assessor – not the lab
    - can enable substantial cost-savings on site development overall
    - minimum UoM not always required – not cost effective
  - Enables better use of on-site measurement tools
    - higher  $U_{opt}$  but cheaper so take more samples – lower overall U.

Ramsey, M. H. (2009). Uncertainty in the assessment of hazard, exposure and risk. *Environmental Geochemistry and Health*, 31, 2, 205-217. doi:10.1007/s10653-008-9211-8



## Acknowledgements

### Co-workers:

- Steve Ellison (Co-author on Eurachem Guide)
- Dr Katy Boon (TPH study in Case Study #2) ,
- Dr Ariadne Argyraki (Case Study #1),
- Dr Paul Taylor (Fitness-for-purpose),
- Dr Jenny Lyn (Sample Prep U)

### Funding:

- EPSRC (Studentship)
- NERC (NE/E009484/1 Environment and Human Health call )
- UK Department of Trade and Industry (now TSB, Partners in Innovation STBF/004/00034C)
- UK Technology Strategy Board (TSB) Technology Programme (Project Number - TP/5/CON/6/I/H0065B).
- CL:AIRE (RP4)
- UK Food Standards Agency (E01034, E01055, E01070)

